



## Research Paper

# Irreversible fluorine covalent organic framework based probe nanoelectrospray ionization mass spectrometry for direct and rapid determination of perfluoroalkyl carboxylic acids

Xu-Qin Ran<sup>a,b</sup>, Shu-Ting Xu<sup>a,b</sup>, Hai-Long Qian<sup>a,b,\*</sup>, Xiu-Ping Yan<sup>a,b,c</sup>

<sup>a</sup> State Key Laboratory of Food Science and Technology, Jiangnan University, Wuxi 214122, China

<sup>b</sup> Institute of Analytical Food Safety, School of Food Science and Technology, Jiangnan University, Wuxi 214122, China

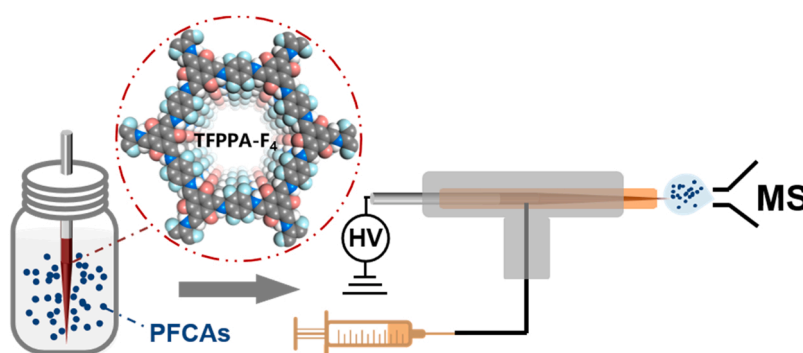
<sup>c</sup> Key Laboratory of Synthetic and Biological Colloids, Ministry of Education, Jiangnan University, Wuxi 214122, China



## HIGHLIGHTS

- Irreversible fluorine COF (FCOF) bonded probe was fabricated via in-situ growth.
- Chemical bonding of irreversible FCOF improved the stability of probe.
- FCOF based PESI-MS was developed for perfluoroalkylcarboxylic acids.
- Mechanism involves hydrophobic, hydrogen bonding and F–F interactions.

## GRAPHICAL ABSTRACT



## ARTICLE INFO

Editor: Joao Pinto da Costa

## Keywords:

Covalent organic framework  
Probe nanoelectrospray ionization mass spectrometry  
Perfluoroalkyl carboxylic acids  
Environmental water  
Extraction

## ABSTRACT

Probe nanoelectrospray ionization mass spectrometry (PESI-MS) is practically desirable for rapid and ultra-sensitive analysis of trace contaminants in environment, but limited with the stable and selective probe coating. Herein, we show the design and preparation of irreversible fluorine-based covalent organic framework (TFPPA-F<sub>4</sub>) covalently bonded probe to couple with ESI-MS (TFPPA-F<sub>4</sub>-PESI-MS) for direct and rapid determination of perfluoroalkyl carboxylic acids (PFCAs) in environmental water. Chemical bonding coating of irreversible crystalline TFPPA-F<sub>4</sub> not only improved stability of the probe, but also offered accessible multiple interactions including hydrophobic, hydrogen bonding and F-F interactions to promote the kinetics and selectivity for PFCAs. The proposed TFPPA-F<sub>4</sub>-PESI-MS realized rapid determination of PFCAs (about 4 min) with low limits of detection of 0.06–0.88 ng L<sup>-1</sup> and wide linear range of 1–5000 ng L<sup>-1</sup> (R<sup>2</sup> of 0.9982–0.9998). Recoveries for the spiked lake and pond water were 85.9–111.1 %. TFPPA-F<sub>4</sub> based probe can maintain the extraction performance after 100 times of extraction. This work shows the great potential of the irreversible covalent organic framework based PESI-MS in rapid and ultra-sensitive determination of contaminants in environmental samples.

\* Corresponding author at: State Key Laboratory of Food Science and Technology, Jiangnan University, Wuxi 214122, China.

E-mail address: [hlqian@jiangnan.edu.cn](mailto:hlqian@jiangnan.edu.cn) (H.-L. Qian).

<https://doi.org/10.1016/j.jhazmat.2023.131584>

Received 12 March 2023; Received in revised form 20 April 2023; Accepted 4 May 2023

Available online 5 May 2023

0304-3894/© 2023 Elsevier B.V. All rights reserved.

## 1. Introduction

Protection of water resource is attracting worldwide attention in coming century due to the irreplaceability of water resource [22,36]. As new emerging contaminant, perfluoroalkyl carboxylic acids (PFCAs), typically represented with perfluorooctanoic acid (PFOA), widely exist in environmental water owing to their extensive application in various industries including aerospace, textile, packaging, automobile, and decoration [28,37]. The abundant C-F bonds with high bond energy make PFCAs extremely stable and difficult to degrade in the environment, resulting in the high bioaccumulation and potential risk for organisms [1,24,5]. Many national regulatory agencies start to monitor PFCAs levels in water (Committee POPRC, 2011; Committee JISC, 2022). A rapid and ultra-sensitive analytical method for PFCAs in environmental water remains indispensable and significant for investigation of the environmental exposure level and toxicological effect of PFCAs.

Trace level and serious interference of complex matrix from real samples make extraction necessary for determination of PFCAs with the primarily performed liquid chromatography-mass spectrometry (LC-MS) [8,30]. However, time-consuming and tedious processes of extraction and separation limit rapid determination of PFCAs with LC-MS [15, 16,38]. Probe nano-electrospray ionization mass spectrometry (PESI-MS) opens up a new stage for the determination of PFCAs [35,3,7]. Different from traditional LC-MS isolated with extraction, PESI-MS can utilize charged solvent to achieve on-line elution of enriched target from the coating of probe to produce spray ions of target for direct MS analysis [23,4,2], offering the opportunity for rapid and ultra-sensitive determination of PFCAs. However, the coatings of commercial probes including polydimethylsiloxane, divinyl-benzene and poly(acrylate) are not particularly selective for the enrichment of PFCAs [34]. Novel efficient coatings of probe for PFCAs are still extremely urgent.

Crystalline porous materials of covalent organic frameworks (COFs), linked with organic monomers by covalent bonds, have already shown up-and-coming talents as adsorbents for sample pretreatment [9,21,26, 27,6]. High stability, ordered porosity, designable structure and tunable functionality allow COFs to rapidly and selectively interact with analytes. Thus, development of COFs based PESI-MS is promising in rapid and ultra-sensitive analysis of complex samples, but still in preliminary stage [10,11,13]. The majority of reported COFs in PESI-MS are reversible, and the COFs are physically adhered on the probe, which greatly compromises the stability and durability of the COFs based probe.

Herein, we intentionally designed to prepare irreversible COF covalently bonded probe to couple with ESI-MS for direct and rapid determination of PFCAs in environmental water. An irreversible crystalline fluorine-based COF (TFPPA-F<sub>4</sub>) condensed with 1,3,5-triformylphloroglucinol (TFP) and 2,3,5,6-tetrafluoro-1,4-phenylenediamine (PA-F<sub>4</sub>) was selected and covalently in-situ grown on stainless steel probe (SS). The highly ordered structure of TFPPA-F<sub>4</sub> with rich F-F, H-bonding and hydrophobic interactions would render the TFPPA-F<sub>4</sub> bonded probe superior sensitivity and rapid kinetics for PFCAs, allowing TFPPA-F<sub>4</sub>-PESI-MS for rapid and ultra-sensitive determination of PFCAs in real environmental water. This work not only expands the talent of irreversible COFs in PESI-MS, but also reveals the great potential of COFs based PESI-MS in rapid and ultra-sensitive determination of contaminants in environmental samples.

## 2. Materials and methods

### 2.1. Materials and chemicals

All chemicals and reagents are commercially available and used without further purification. TFP was obtained from Yanshen

Technology Co., Ltd. (Jilin, China). PA-F<sub>4</sub> was bought from Tokoy Chemical Industry Co., Ltd. (Shanghai, China). 1,4-Dioxane, mesitylene, acetic acid (HAc), tetrahydrofuran (THF), (3-aminopropyl)triethoxysilane (APTES), perfluorohexanoic acid (PFHxA), perfluoroheptanoic acid (PFHpA), PFOA, perfluorononanoic acid (PFNA), perfluorodecanoic acid (PFDA) and perfluoroundecanoic acid (PFUnDA) were purchased from Aladdin Chemistry Co., Ltd. (Shanghai, China). Perfluoro-n-(1,2,3,4-<sup>13</sup>C<sub>4</sub>) octanoic acid (<sup>13</sup>C<sub>4</sub>-PFOA) as the internal standard compound (IS) was obtained from Wellington laboratories Inc. (Guelph (ON), Canada). HPLC grade methanol (MeOH) was gotten from Fisher Chemical (Shanghai, China). Ultrapure water was purchased from Wahaha Foods Co., Ltd. (Shanghai, China). SS (diameter ca. 0.25 mm, length 40 mm) was purchased from Huaer Medical Instrument Co., Ltd (Hebei, China). Fused silica capillary (o.d. 0.53 mm) was purchased from Yongnian Optic Fiber Plant (Hebei, China). Environmental water samples were collected locally and directly used as matrix samples without any pretreatment.

### 2.2. Instrumentation

Powder X-ray diffraction (PXRD) patterns were recorded on a D2 PHASER X-ray diffractometer (Bruker, German) using Cu K $\alpha$  radiation ( $\lambda = 1.5418 \text{ \AA}$ ) with a scanning speed of  $8^\circ \text{ min}^{-1}$  and a step size of  $0.05^\circ$  in  $2\theta$ . Scanning electron microscopy (SEM) images were recorded on an S-3500 N (Hitachi, Japan) scanning electron microscope. Fourier transform infrared (FTIR) spectra were measured on a Nicolet IR IS10 spectrometer (Nicolet, USA). X-ray photoelectron spectroscopy (XPS) experiments were performed on Axis supra (Kratos, UK). N<sub>2</sub> adsorption experiments were performed on Autosorb-iQ (Quantachrome, USA) using N<sub>2</sub> adsorption at 77 K. Mass spectra were obtained on QTRAP 4500 mass spectrometer (AB SCIEX, USA) equipped with a Nanospray II Source (P/N #1004600, AB SCIEX) and analyst software (AB SCIEX, USA) to control equipment and acquire data.

### 2.3. Preparation of irreversible TFPPA-F<sub>4</sub> bonded probe

SS was dipped into aqua regia (depth ca. 1.5 cm) for 1 min, then washed with ultrapure water continuously until pH reached 7.0, and finally dried at room temperature. The dried etched SS was immersed into a mixture of APTES/MeOH/H<sub>2</sub>O (v/v/v, 4/5/1) solution for 4 h and dried in a vacuum oven at 120 °C for 1 h to obtain the amino-functionalized SS (NH<sub>2</sub>-SS). The NH<sub>2</sub>-SS was immersed in the mixture of TFP (31.5 mg), PA-F<sub>4</sub> (40.5 mg), 6 M HAc (0.1 mL), 1,4-dioxane (0.5 mL), and mesitylene (0.5 mL) with ultrasonic process for 10 min, degassed with freeze-pump-thaw cycles, reacted at 100 °C for 3 days, washed with THF and finally dried in a vacuum oven for 2 h to obtain TFPPA-F<sub>4</sub> bonded SS (TFPPA-F<sub>4</sub>-SS).

### 2.4. Procedure of TFPPA-F<sub>4</sub>-PESI-MS

All the samples were extracted with TFPPA-F<sub>4</sub>-SS. Typically, the TFPPA-F<sub>4</sub>-SS was first cleaned with MeOH and ultrapure water, immersed into 1 mL of IS (<sup>13</sup>C<sub>4</sub>-PFOA) mixed samples with vigorous stirring at room temperature for 2 min, then washed with ultrapure water for 10 s, and finally moved to a home-made probe holder (Fig. S1). The home-made probe holder with a three-dimensional moving platform and hand-held digital microscope can finely control and monitor the position of the TFPPA-F<sub>4</sub>-SS (Fig. 1a). The distance between tip of TFPPA-F<sub>4</sub>-SS and MS inlet was controlled as ca. 5 mm. The prepared TFPPA-F<sub>4</sub>-SS also served as ESI emitter. A negative voltage of 3.2 kV, supplied by the MS-integrated high-voltage source, was applied to the end of COF-probe (Fig. 1b). An MS-integrated syringe pump was used to offer solvent as eluent. The eluent of MeOH at a flow rate of  $20 \mu\text{L min}^{-1}$  was pumped to elute the TFPPA-F<sub>4</sub> coating on the probe. The produced

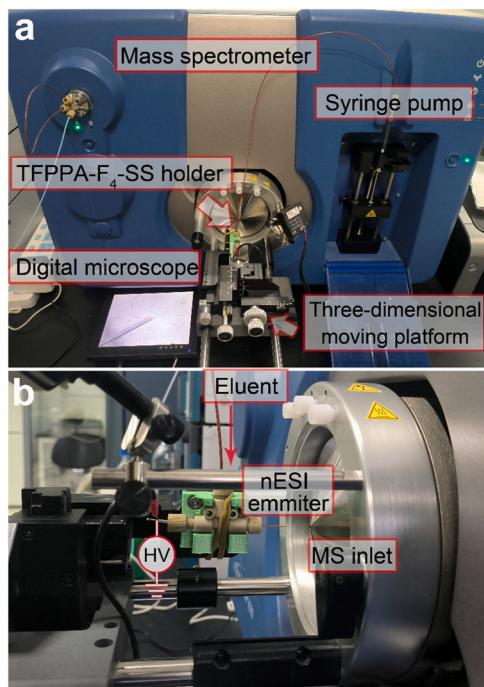
charged spray of eluate was directly introduced into MS for data acquisition and analysis. The TFPPA-F<sub>4</sub>-SS could be recycled by vigorous stirring with MeOH and water.

### 3. Results and discussion

#### 3.1. Preparation and characterization of irreversible TFPPA-F<sub>4</sub> bonded probe

Hydrophobic, hydrogen bonding, and F-F interaction are well-known for promoting the selectivity of adsorbent for PFCAs [13,14,19]. Ordered crystalline structure of adsorbents can facilitate mass transfer of analytes to achieve rapid adsorption [18,29]. Accordingly, irreversible crystalline COF TFPPA-F<sub>4</sub> with rich hydrophobic aromatic rings, H-bond acceptors (C=O), H-bond donors (N-H) and F groups was selected as the coating to promote the selectivity and kinetics of probe to PFCAs. The TFPPA-F<sub>4</sub> covalently bonded probe was synthesized via in-situ growth approach as shown in Fig. 2. The etched SS was first reacted with APTES to prepare NH<sub>2</sub>-SS. Then, TFP and PA-F<sub>4</sub> were further introduced to react with NH<sub>2</sub>-SS via the Schiff-base reaction to obtained TFPPA-F<sub>4</sub>-SS.

Formation of TFPPA-F<sub>4</sub> and its bonded SS was first confirmed with PXRD. The prepared TFPPA-F<sub>4</sub> and TFPPA-F<sub>4</sub> scraped from TFPPA-F<sub>4</sub>-SS gave an evident peak at 4.7° in PXRD pattern, which was in good consistent with the simulated PXRD pattern of TFPPA-F<sub>4</sub>. The simulation and refinement gave the specific cell parameters of space group P6/m,  $a = b = 21.8234 \text{ \AA}$ ,  $c = 3.5031 \text{ \AA}$ ,  $\alpha = \beta = 90^\circ$  and  $\gamma = 120^\circ$  for TFPPA-F<sub>4</sub>



**Fig. 1.** (a) Photograph of TFPPA-F<sub>4</sub>-PESI-MS system with home-made probe holder, three-dimensional fine moving platform, hand-held digital microscope, syringe pump, and mass spectrometer. (b) Detailed fittings including the home-made probe holder, TFPPA-F<sub>4</sub>-SS served as nanoelectrospray ionization (nESI) emitter, MS-integrated high-voltage source and inlet of MS.

(Fig. 3a-c and Table S1). Irreversible COFs are famous for their superior stability [12,25,32,33]. The irreversible keto-enamine COFs can keep the crystallinity even in extreme acid/base condition [17]. Hydrolytic stability of TFPPA-F<sub>4</sub> was examined with the PXRD. No obvious change of the PXRD patterns for TFPPA-F<sub>4</sub> after immersing in different solvents demonstrated the high chemical stability of TFPPA-F<sub>4</sub> coating in 1 M NaOH/HCl, water and MeOH (Fig. 3d). In contrast, the reported reversible COF-F-1 for the PFCAs was totally hydrolyzed in 1 M HCl and lost the crystallinity in 1 M NaOH, MeOH and H<sub>2</sub>O, indicating the practical potential of irreversible COFs in complex matrix (Fig. S2).

Change and formation of chemical bonds were verified with FTIR. The FTIR spectra of TFPPA-F<sub>4</sub> appeared a new shoulder peak at 1662 cm<sup>-1</sup> (assigned to the stretching band of C=O), which overlapped with the strong band of C=C (1593 cm<sup>-1</sup>), indicating formation of irreversible keto structure. Intense stretching peak at 1010 cm<sup>-1</sup> revealed abundant C-F in the prepared TFPPA-F<sub>4</sub> (Fig. 3e). Evident peak of F 1s in wide XPS scanning spectra also conveyed the existence of rich F in the TFPPA-F<sub>4</sub> (Fig. 3f).

Optical photograph of TFPPA-F<sub>4</sub>-SS showed evident orange red TFPPA-F<sub>4</sub> coating (about 1.5 cm) on the end of SS (Fig. 4a). SEM image revealed the sphere-like morphology of TFPPA-F<sub>4</sub> (Fig. 4b). Compared to NH<sub>2</sub>-SS, TFPPA-F<sub>4</sub>-SS offered larger density of TFPPA-F<sub>4</sub> particles on surface of SS (Fig. 4c and d), further confirming successful bonding of TFPPA-F<sub>4</sub> with SS. SEM image of cross-section for TFPPA-F<sub>4</sub>-SS reveals about 25 μm thickness of TFPPA-F<sub>4</sub> (Fig. 4e).

N<sub>2</sub> adsorption experiment characterized the pore properties of the prepared irreversible TFPPA-F<sub>4</sub> with the BET surface area of 233 m<sup>2</sup> g<sup>-1</sup> and narrow pore size distribution of 1.0–3.0 nm (Fig. S3 and S4). The main pores were distributed at ~1.79 nm. The large surface area and suitable pore size of TFPPA-F<sub>4</sub> with high chemical stability as well as rich aromatic ring, C=O, N-H and F groups make selective extraction of PFCAs with TFPPA-F<sub>4</sub> reasonable, inspiring us to explore the potential of TFPPA-F<sub>4</sub> based PESI-MS for the analysis of PFCAs.

#### 3.2. Development of TFPPA-F<sub>4</sub>-PESI-MS

TFPPA-F<sub>4</sub>-SS was immersed into 1 mL of samples containing six PFCAs of PFHxA, PFHpA, PFOA, PFNA, PFDA, and PFUnDA. The PFCAs extracted TFPPA-F<sub>4</sub>-SS (TFPPA-F<sub>4</sub>-SS-PFCAs) was washed with water, then equipped on a home-made probe holder and linked with MS (Fig. 1 and S1). Eluent of MeOH was set at a flow rate of 20 μL min<sup>-1</sup> under a negative voltage of 3.2 kV to elute the extracted PFCAs to produce charged spray of PFCAs for direct MS analysis. In order to investigate extraction time, signal intensities of PFCAs ( $I_{\text{analyte}}$ ) was monitored at different time. It was found that the  $I_{\text{analyte}}$  remained no evident change after 2 min (Fig. S5). Thus, the extraction time was considered to be 2 min. Elution profile showed complete elution of PFCAs from TFPPA-F<sub>4</sub> based probe in 1.5 min (Fig. 5a).

Twenty precursor ion/product ion pairs of the six PFCAs and <sup>13</sup>C<sub>4</sub>-PFOA (IS) were monitored in MRM model to qualitatively and quantitatively analyze PFCAs under the optimal conditions in Table S2, Fig. 5 and Fig. S6. Quantitative performance of the established TFPPA-F<sub>4</sub>-PESI-MS method was assessed via analyzing pure water samples spiked with 0.5 – 5000 ng L<sup>-1</sup> of PFCAs and 100 ng L<sup>-1</sup> of <sup>13</sup>C<sub>4</sub>-PFOA as IS. Ratio of signal intensities for PFCAs and IS ( $I_{\text{analyte}}/I_{\text{IS}}$ ) was used for construction of calibration curve and quantitative calculation. The  $I_{\text{analyte}}/I_{\text{IS}}$  linearly increased with the concentration of PFCAs from 5 to 5000 ng L<sup>-1</sup> except of PFNA and PFDA (1–5000 ng L<sup>-1</sup>) with determination coefficients ( $R^2$ ) of 0.9982–0.9998 (Table 1 and Fig. S7). Limits of detection (LODs, S/N = 3) and quantification (LOQs, S/N = 10) of TFPPA-F<sub>4</sub>-PESI-MS for the PFCAs were 0.06–0.88 and 0.20–2.92 ng L<sup>-1</sup>, respectively (Table 1).

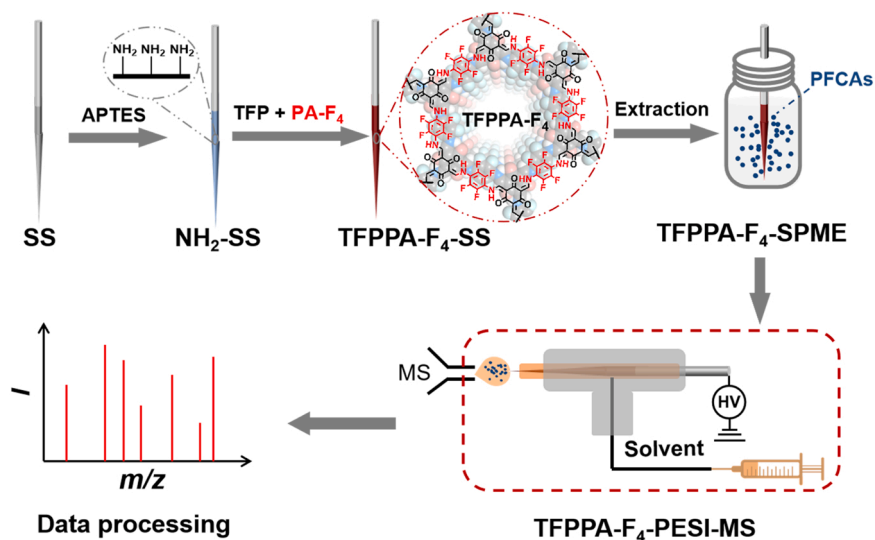


Fig. 2. Schematic illustration of the preparation of TFPPA-F<sub>4</sub>-SS for PESI-MS analysis of PFCAs.

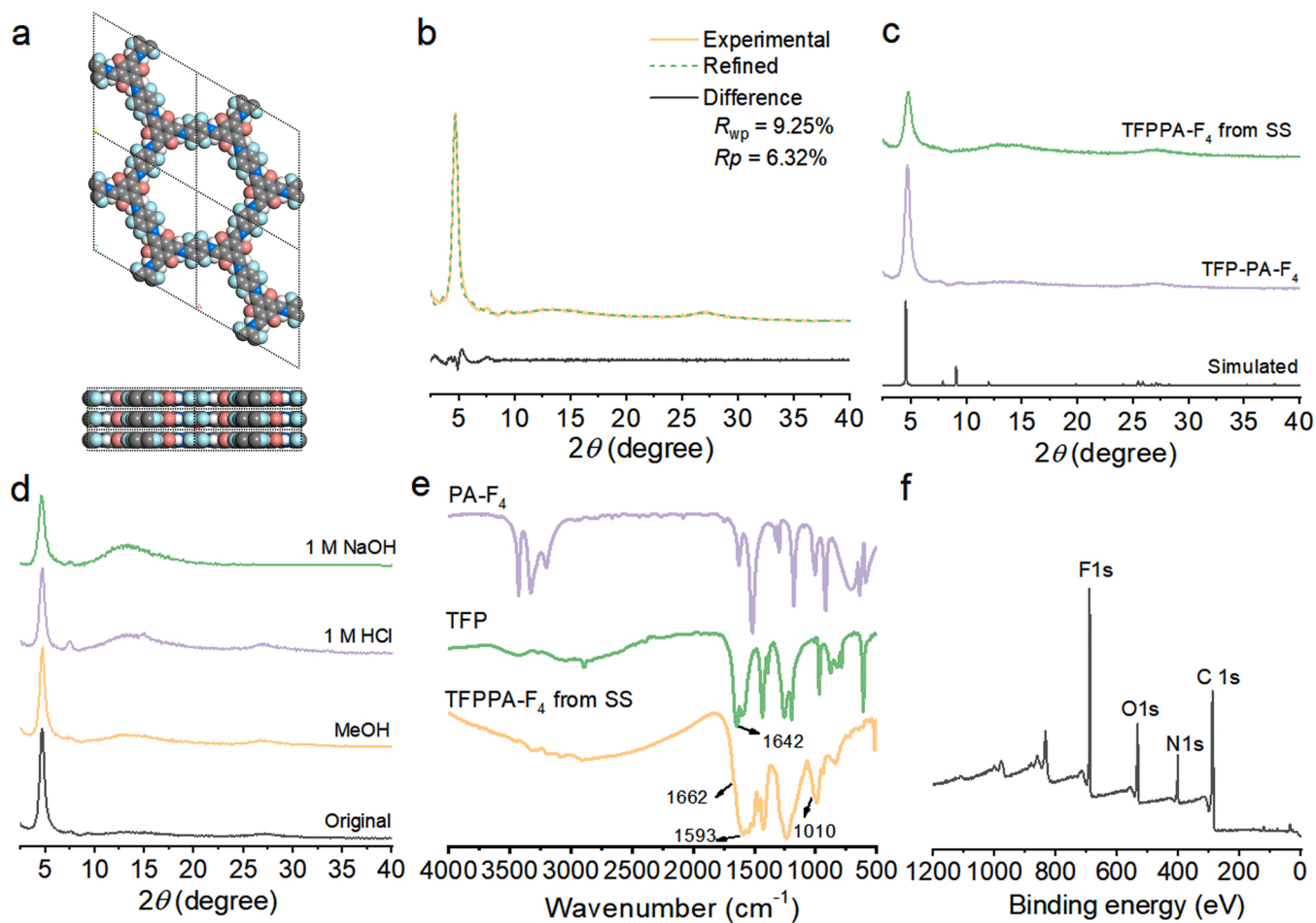
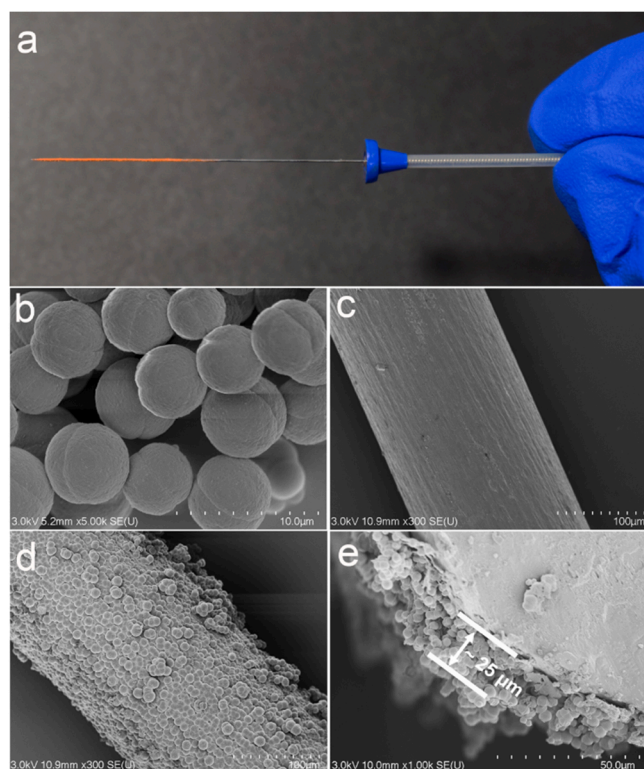
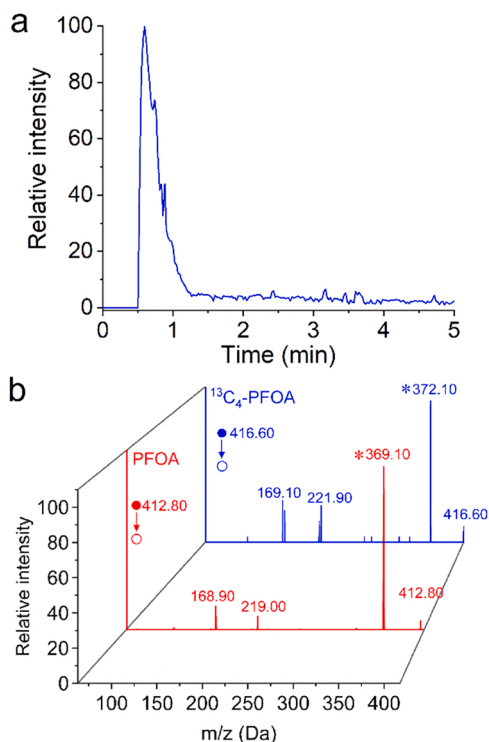


Fig. 3. (a) Unit cell of TFPPA-F<sub>4</sub>. (b) Pawley refinement of TFPPA-F<sub>4</sub>. (c) Experimental and simulated PXRD patterns of TFPPA-F<sub>4</sub>. (d) PXRD patterns of TFPPA-F<sub>4</sub> treated with different solvents. (e) FTIR spectra of TFP, PA-F<sub>4</sub> and TFPPA-F<sub>4</sub>. (f) Wide XPS scanning spectra of TFPPA-F<sub>4</sub>.





**Fig. 4.** (a) Optical photograph of TFPPA-F<sub>4</sub>-SS. SEM images of (b) TFPPA-F<sub>4</sub>, (c) etched SS, (d) TFPPA-F<sub>4</sub>-SS and (e) cross-section for the TFPPA-F<sub>4</sub>-SS.



**Fig. 5.** (a) Elution profile for PFCAs (5 μg L<sup>-1</sup>). (b) MS/MS spectra of PFOA and <sup>13</sup>C<sub>4</sub>-PFOA (The ions labeled with star were selected for quantitation).

Moreover, the developed TFPPA-F<sub>4</sub>-PESI-MS not only did not require separation process of LC, but also facilitated extraction speed for PFCAs because of ordered crystalline structure of TFPPA-F<sub>4</sub>, resulting in shorter analysis time of PFCAs (about 4 min) in contrast to traditional LC-MS. Compared with many other reported analytical methods for PFCAs, our proposed TFPPA-F<sub>4</sub>-PESI-MS was quite competitive in sample throughput and sensitivity (Table S3).

Maximum molecular sizes of PFHxA-PFUnDA were calculated to be smaller than 14.9 Å (Fig. S8), which was covered by the pore size of TFPPA-F<sub>4</sub>, favoring size exclusion effect and pore selection. XPS characterization was further applied to investigate the effect of TFPPA-F<sub>4</sub> on the extraction of PFCAs. The results showed O 1s, N 1s and F 1s spectra of TFPPA-F<sub>4</sub> changed after adsorption of PFCAs. XPS peaks of TFPPA-F<sub>4</sub> for N 1s at 404.20 eV, O 1s at 532.83 eV and F 1s at 691.32 eV likely resulted from the intramolecular hydrogen bonding. After the adsorption of PFCAs, binding energies of the aforementioned N 1s, O 1s and F 1s shifted to 403.53, 533.04 and 690.46 eV, respectively (Fig. S9). These changes indicate the formation of hydrogen bonding and F-F interaction of TFPPA-F<sub>4</sub> and PFCAs [20,31].

Repeatability and reproducibility are crucial for probe. Relative standard deviations (RSDs) of  $I_{\text{analyte}}/I_{\text{IS}}$  for 12 replicate extractions of PFCAs with TFPPA-F<sub>4</sub>-SS were 1.87–4.6 %. The RSDs of  $I_{\text{analyte}}/I_{\text{IS}}$  for PFCAs with three different TFPPA-F<sub>4</sub>-SS were 6.7–12.5 % (Table 1). The  $I_{\text{analyte}}/I_{\text{IS}}$  of 100 ng L<sup>-1</sup> PFCAs gave no significant change after 100 extraction cycles, indicating long-lasting reusability of the prepared TFPPA-F<sub>4</sub> bonded probe (Fig. S10). The stability of TFPPA-F<sub>4</sub> coating was also verified by PXRD and FTIR. Compared with original TFPPA-F<sub>4</sub> coating, the TFPPA-F<sub>4</sub> coating after extraction gave no obvious change in FTIR spectra and PXRD patterns (Fig. S11 and S12), indicating good retention and no hydrolysis of the crystalline structure of TFPPA-F<sub>4</sub> coating after extraction, which was attribute to the irreversibility of TFPPA-F<sub>4</sub> as well as the chemical bonding of TFPPA-F<sub>4</sub> and SS.

### 3.3. TFPPA-F<sub>4</sub>-PESI-MS for real samples

Practicability of the developed TFPPA-F<sub>4</sub>-PESI-MS method was verified by analyzing environmental water samples. Matrix factor (MF) of three environmental water samples was 85.7–114.7 % (Fig. S13), indicating that the standard calibration curves obtained with pure water are available for quantitative calculation of real samples. PFCAs were found in all the three samples. In lake water I, PFHxA, PFHpA, PFOA, PFNA and PFDA were found to be  $9.4 \pm 1.2$ ,  $19.0 \pm 0.2$ ,  $4.5 \pm 0.2$ ,  $18.2 \pm 0.2$ , and  $10.9 \pm 0.2$  ng L<sup>-1</sup>, respectively. In lake water II, PFOA, PFNA and PFDA were detected to be  $3.4 \pm 0.2$ ,  $8.3 \pm 0.4$ , and  $2.4 \pm 0.2$  ng L<sup>-1</sup>, respectively. In pond water, PFHxA, PFHpA, PFOA, PFNA and PFDA was found to be  $20.3 \pm 1.6$ ,  $27.1 \pm 1.0$ ,  $4.9 \pm 0.6$ ,  $8.5 \pm 0.7$ , and  $3.5 \pm 0.5$  ng L<sup>-1</sup>, respectively (Table S4). Recoveries of spiked 100 ng L<sup>-1</sup> of PFCAs in the three water samples were 82.9–111.1 % (Table S4), indicating good accuracy of the TFPPA-F<sub>4</sub>-PESI-MS for direct determination of trace level PFCAs in real samples.

## 4. Conclusions

In summary, an irreversible fluorine-based COF TFPPA-F<sub>4</sub> bonded probe coupled with ESI-MS was firstly designed for determination of PFCAs. The rich hydrophobic aromatic rings, H-bond acceptors/donors and F groups of TFPPA-F<sub>4</sub> provided the conjunction of hydrophobic, hydrogen bonding, and F-F interactions to promote the selectivity of TFPPA-F<sub>4</sub> for PFCAs. The irreversible TFPPA-F<sub>4</sub> with ordered structure and great stability further served as probe coating via chemical bonding method to render the TFPPA-F<sub>4</sub> based probe of PFCAs with fast kinetics

**Table 1**  
Validation of TFPPA-F<sub>4</sub>-PESI-MS for PFCAs.

Analyte	Linear range (ng L <sup>-1</sup> )	Regression equation	R <sup>2</sup>	LOD (ng L <sup>-1</sup> )	LOQ (ng L <sup>-1</sup> )	Repeatability (RSD, %) <sup>a</sup>	
						One probe (n = 6)	Probe-to-probe (n = 3)
PFHxA	5–5000	Y = 0.0038X + 0.0716	0.9984	0.88	2.92	4.6	6.7
PFHpA	5–5000	Y = 0.0161X + 0.2132	0.9998	0.78	2.60	2.1	7.5
PFOA	5–5000	Y = 0.0619X + 0.5404	0.9996	0.26	0.87	2.2	10.9
PFNA	1–5000	Y = 0.0260X + 0.2428	0.9982	0.41	1.35	1.9	8.6
PFDA	1–5000	Y = 0.0728X + 0.7818	0.9996	0.30	0.99	4.1	6.8
PFUnDA	5–5000	Y = 0.2310X + 22.517	0.9991	0.06	0.20	2.5	12.5

<sup>a</sup> 100 ng L<sup>-1</sup> of PFCAs.

and high selectivity. Coupled with ESI-MS, the proposed TFPPA-F<sub>4</sub>-PESI-MS achieved rapid and ultra-sensitive analysis of trace PFCAs in environmental samples. Moreover, the chemical bonding coating improved the stability of the COF based probe. This work not only offers an efficient analytical method for trace PFCAs, but also reveals the great potential of irreversible COFs-PESI-MS technology in rapid and accurate determination of pollutants in environmental samples.

### Environmental Implication

As new emerging contaminant, perfluoroalkyl carboxylic acids (PFCAs) widely exist in environmental water. Compared with the traditional LC-MS, probe nanoelectrospray ionization mass spectrometry (PESI-MS) opens up a new stage for rapid and sensitive determination of PFCAs owing to the high selectivity and no LC separation, but is limited by the efficient coating of probe. Here, an irreversible fluorine-based covalent organic framework (TFPPA-F<sub>4</sub>) covalently bonded probe coupled with ESI-MS (TFPPA-F<sub>4</sub>-PESI-MS) was designed and prepared for direct and rapid determination of PFCAs in environmental water.

### CRedit authorship contribution statement

**Xu-Qin Ran:** Conceptualization, Investigation, Methodology, Data curation, Writing – original draft. **Shu-Ting Xu:** Methodology, Resources. **Hai-Long Qian:** Conceptualization, Methodology, Writing – review & editing, Supervision, Funding acquisition. **Xiu-Ping Yan:** Writing – review & editing, Funding acquisition.

### Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

### Data Availability

Data will be made available on request.

### Acknowledgment

This study is supported by the National Natural Science Foundation of China (Nos. 22076066 and 22176073), the Program of “Collaborative Innovation Center of Food Safety and Quality Control in Jiangsu Province” and the “Fundamental Research Funds for the Central Universities”.

### Appendix A. Supporting information

Supplementary data associated with this article can be found in the online version at [doi:10.1016/j.jhazmat.2023.131584](https://doi.org/10.1016/j.jhazmat.2023.131584).

### References

- [1] Ateia, M., Alsaiee, A., Karanfil, T., Dichtel, W., 2019. Efficient PFAS removal by amine-functionalized sorbents: critical review of the current literature. *Environ. Sci Tech Lett* 6, 688–695.
- [2] Deng, J., Yang, Y., Wang, X., Luan, T., 2014. Strategies for coupling solid-phase microextraction with mass spectrometry. *TrAC-Trend Anal Chem* 55, 55–67.
- [3] Deng, J., Yang, Y., Fang, L., Lin, L., Zhou, H., Luan, T., 2014. Coupling solid-phase microextraction with ambient mass spectrometry using surface coated wooden-tip probe for rapid analysis of ultra trace perfluorinated compounds in complex samples. *Anal Chem* 86, 11159–11166.
- [4] Deng, J., Yang, Y., Liu, Y., Fang, L., Lin, L., Luan, T., 2019. Coupling paternò-büchi reaction with surface-coated probe nanoelectrospray ionization mass spectrometry for in vivo and microscale profiling of lipid C=C location isomers in complex biological tissues. *Anal Chem* 91, 4592–4599.
- [5] DeWitt, J.C., Peden-Adams, M.M., Keller, J.M., Germolec, D.R., 2012. Immunotoxicity of perfluorinated compounds: recent developments. *Toxicol Pathol* 40, 300–311.
- [6] Ding, S.Y., Wang, W., 2013. Covalent organic frameworks (COFs): from design to applications. *Chem Soc Rev* 42, 548–568.
- [7] Fang, L., Deng, J., Yang, Y., Wang, X., Chen, B., Liu, H., et al., 2016. Coupling solid-phase microextraction with ambient mass spectrometry: strategies and applications. *TrAC-Trend Anal Chem* 85, 61–72.
- [8] Feng, H., Ruan, Y., Zhang, K., Lam, P.K.S., 2019. Current analytical methodologies and gaps for per- and polyfluoroalkyl substances determination in the marine environment. *TrAC-Trend Anal Chem* 121, 115372.
- [9] Feng, J., Feng, J., Ji, X., Li, C., Han, S., Sun, H., et al., 2021. Recent advances of covalent organic frameworks for solid-phase microextraction. *TrAC-Trend Anal Chem* 137, 116208.
- [10] Gao, W., Cheng, J., Yuan, X., Tian, Y., 2021. Covalent organic framework-graphene oxide composite: A superior adsorption material for solid phase microextraction of bisphenol A. *Talanta* 222, 121501.
- [11] Gao, W., Li, G., Liu, H., Tian, Y., Li, W.T., Fa, Y., et al., 2021. Covalent organic frameworks with tunable pore sizes enhanced solid-phase microextraction direct ionization mass spectrometry for ultrasensitive and rapid analysis of tetrabromobisphenol A derivatives. *Sci Total Environ* 764, 144388.
- [12] Guan, X., Li, H., Ma, Y., Xue, M., Fang, Q., Yan, Y., et al., 2019. Chemically stable polyarylether-based covalent organic frameworks. *Nat Chem* 11, 587–594.
- [13] Hou, Y.J., Deng, J., He, K., Chen, C., Yang, Y., 2020. Covalent organic frameworks-based solid-phase microextraction probe for rapid and ultrasensitive analysis of trace per- and polyfluoroalkyl substances using mass spectrometry. *Anal Chem* 92, 10213–10217.
- [14] Huang, J., Shi, Y., Huang, G., Huang, S., Zheng, J., Xu, J., et al., 2022. Facile synthesis of a fluorinated-squaramide covalent organic framework for the highly efficient and broad-spectrum removal of per- and polyfluoroalkyl pollutants. *Angew Chem Int Ed* 61, e202206749.
- [15] Huang, Y., Zhang, W., Bai, M., Huang, X., 2020. One-pot fabrication of magnetic fluorinated carbon nanotubes adsorbent for efficient extraction of perfluoroalkyl carboxylic acids and perfluoroalkyl sulfonic acids in environmental water samples. *Chem Eng J* 380, 122392.
- [16] Jia, Y., Qian, J., Pan, B., 2021. Dual-functionalized MIL-101(Cr) for the selective enrichment and ultrasensitive analysis of trace per- and poly-fluoroalkyl substances. *Anal Chem* 93, 11116–11122.
- [17] Kandambeth, S., Mallick, A., Lukose, B., Mane, M.V., Heine, T., Banerjee, R., 2012. Construction of crystalline 2D covalent organic frameworks with remarkable chemical (acid/base) stability via a combined reversible and irreversible route. *J Am Chem Soc* 134, 19524–19527.
- [18] Karak, S., Dey, K., Torris, A., Halder, A., Bera, S., Kanheerampokil, F., et al., 2019. Inducing disorder in order: hierarchically porous covalent organic framework nanostructures for rapid removal of persistent organic pollutants. *J Am Chem Soc* 141, 7572–7581.
- [19] Koda, Y., Terashima, T., Sawamoto, M., 2014. Fluorous microgel star polymers: selective recognition and separation of polyfluorinated surfactants and compounds in water. *J Am Chem Soc* 136, 15742–15748.
- [20] Li, X., Wang, Y., Wen, J., Zheng, L., Qian, C., Cheng, Z., et al., 2023. Porous organic polycarbene nanotrap for efficient and selective gold stripping from electronic waste. *Nat Commun* 14, 263.

- [21] Liu, R., Tan, K.T., Gong, Y., Chen, Y., Li, Z., Xie, S., et al., 2021. Covalent organic frameworks: an ideal platform for designing ordered materials and advanced applications. *Chem Soc Rev* 50, 120–242.
- [22] Makanda, K., Nzama, S., Kanyerere, T., 2022. Assessing the role of water resources protection practice for sustainable water resources management: a review. *Water* 14, 3153.
- [23] Mirabelli, M.F., Wolf, J.C., Zenobi, R., 2016. Direct coupling of solid-phase microextraction with mass spectrometry: sub-pg/g sensitivity achieved using a dielectric barrier discharge ionization source. *Anal Chem* 88, 7252–7258.
- [24] Nguyen, G., Nocentini, A., Angeli, A., Gratteri, P., Supuran, C., Donald, W., 2020. Perfluoroalkyl substances of significant environmental concern can strongly inhibit human carbonic anhydrase isozymes. *Anal Chem* 92, 4614–4622.
- [25] Qian, H.L., Meng, F.L., Yang, C.X., Yan, X.P., 2020. Irreversible amide-linked covalent organic framework for selective and ultrafast gold Recovery. *Angew Chem Int Ed* 59, 17607–17613.
- [26] Qian, H.-L., Wang, Y., Yan, X.-P., 2022. Covalent organic frameworks for environmental analysis. *TrAC-Trend Anal Chem* 147, 116516.
- [27] Qiu, S., Fang, Q., Li, H., Guan, X., Yusran, Y., 2020. Postsynthetic functionalization of covalent organic frameworks. *Nat Sci Rev* 7, 170–190.
- [28] Rahman, M., Peldszus, S., Anderson, W., 2014. Behaviour and fate of perfluoroalkyl and polyfluoroalkyl substances (PFASs) in drinking water treatment: a review. *Water Res* 50, 318–340.
- [29] Ran, X.-Q., Qian, H.-L., Yan, X.-P., 2022. Integrating ordered two-dimensional covalent organic frameworks to solid-state nanofluidic channels for ultrafast and sensitive detection of mercury. *Anal Chem* 94, 8533–8538.
- [30] Ruan, T., Jiang, G., 2017. Analytical methodology for identification of novel per- and polyfluoroalkyl substances in the environment. *TrAC-Trend Anal Chem* 95, 122–131.
- [31] Stevens, J.S., Byard, S.J., Seaton, C.C., Sadiq, G., Davey, R.J., Schroeder, S.L., 2014. Proton transfer and hydrogen bonding in the organic solid state: a combined XRD/XPS/ssNMR study of 17 organic acid-base complexes. *Phys Chem Chem Phys* 16, 1150–1160.
- [32] Stewart, D., Antypov, D., Dyer, M., Pitcher, M., Katsoulidis, A., Chater, P., et al., 2017. Stable and ordered amide frameworks synthesised under reversible conditions which facilitate error checking. *Nat Commun* 8, 1102.
- [33] Su, Y., Wan, Y., Xu, H., Otake, K., Tang, X., Huang, L., et al., 2020. Crystalline and stable benzofuran-linked covalent organic frameworks from irreversible cascade reactions. *J Am Chem Soc* 142, 13316–13321.
- [34] Suwannakot, P., Lisi, F., Ahmed, E., Liang, K., Babarao, R., Gooding, J., et al., 2020. Metal-organic framework-enhanced solid-phase microextraction mass spectrometry for the direct and rapid detection of perfluorooctanoic acid in environmental water samples. *Anal Chem* 92, 6900–6908.
- [35] Suwannakot, P., Lisi, F., Ahmed, E., Liang, K., Babarao, R., Gooding, J., et al., 2020. Metal-organic framework-enhanced solid-phase microextraction mass spectrometry for the direct and rapid detection of perfluorooctanoic acid in environmental water samples. *Anal Chem* 92, 6900–6908.
- [36] Tong, X., Mohapatra, S., Zhang, J., Tran, N., You, L., He, Y., et al., 2022. Source, fate, transport and modelling of selected emerging contaminants in the aquatic environment: Current status and future perspectives. *Water Res* 217, 118418.
- [37] Wang, C., Li, W., Lv, Y., Bai, H., Zhao, P., Wen, L., et al., 2020. Rapid analysis of perfluorinated carboxylic acids in textiles by dielectric barrier discharge ionization-mass spectrometry. *Microchem J* 155, 104773.
- [38] Yao, X., Zhou, Z., He, M., Chen, B., Liang, Y., Hu, B., 2018. One-pot polymerization of monolith coated stir bar for high efficient sorptive extraction of perfluoroalkyl acids from environmental water samples followed by high performance liquid chromatography-electrospray tandem mass spectrometry detection. *J Chromatogr A* 1553, 7–15.